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CO₂ReMoVe

Possible impacts of captured CO₂ stream impurities on transport infrastructure and geological storage formations

Current understanding and implications for EU legislation

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Abstract

At present, European regulation does not specify in quantitative terms a legal composition of a captured CO₂ stream. This report aims to provide an overview of the current state regulation regarding the presence of impurities in captured CO₂ streams, and their potential impacts on transport infrastructure and storage formations. It concludes that although impurities in the CO₂ stream may have impacts on storage efficiency and the porosity of storage reservoirs, if deemed necessary, advice on setting quantitative limits on the presence of a number of impurities for the purposes of safe and efficient storage cannot yet be given. A potential best practice emerging from literature regarding storage capacity could include that the amount of non-condensable gases in the CO₂ stream should not exceed 4% by volume. This figure is understood to reflect an optimum balance between gas conditioning costs and the costs of compression. Further work is recommended that assesses the interaction between the CO₂ purity requirements between the different stages of the CCS chain.

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Summary

Demonstrating the safe and permanent storage of CO₂ in geological storage complexes is important for realising CO₂ capture and storage (CCS). Ongoing research is targeted towards improving understanding of storage integrity issues, focusing on migration processes of the injected CO₂, wellhead integrity, fault reactivation and the occurrence of surface deformations. In addition, the chemical composition of the CO₂ stream can potentially affect the chemistry in the geological storage reservoir. This report aims to provide an overview of the current state regulation regarding the presence of impurities in captured CO₂ streams, and their potential impacts on transport infrastructure and storage formations.

CO₂ streams from all capture processes will contain variable levels of impurities. It is possible to remove such impurities using gas cleaning techniques, however this will increase the overall cost of CCS projects. If left untreated, certain impurities can to some extent change the physical behaviour of the bulk gas, which need to be taken into account in the design of the compression and transport system. Furthermore, there has been a limited amount of research conducted on the potential effects of impurities on storage efficiency and storage integrity.

At present, European regulation does not set quantitative limits on the composition of a captured CO₂ stream, but places the responsibility of the competent authority to determine an acceptable CO₂ stream composition on a case-by-case basis. Although the impacts of impurities on the transport system are relatively well understood, and health and safety issues can be overcome through risk assessment, the limited understanding on the impacts of impurities on storage integrity may cause difficulties in assessing the compositional requirements and laying those down in legislation.

'Impurities', 'contaminants' or 'other components' in CO₂ streams can include nitrogen (N₂), oxygen (O₂) and water (H₂O), but also air pollutants such as sulphur and nitrogen oxides (SO_x and NO_x), particulates, hydrochloric acid (HCl), hydrogen fluorides (HF), mercury, other metals and trace organic and inorganic contaminants. The types of impurities that may be present in a CO₂ stream destined for CO₂ storage depends on fuel type, the nature of the process and the application of capture and separation technologies. For instance, oxyfuel processes may lead to a certain oxygen concentration in the CO₂ stream.

Literature indicates that impurities in the CO₂ stream could have an impact on storage efficiency and the porosity of storage reservoirs. Sulphur dioxides (SO_x) in the CO₂ stream, and the subsequent formation of sulphuric acids in the well appear to lead to increased geochemical activity over pure CO₂. Effects could go two ways: SO_x could accelerate mineralisation of CO₂ reducing the possibility of CO₂ leakage, but it has also been argued that it could negatively affect the cap rock and well closing infrastructure. The role of other impurities is even less clear at this point. A best practice that is suggested by literature regarding storage capacity could include that the amount of non-condensable gases in the CO₂ stream should not exceed 4% by volume, reflecting a balance between gas conditioning costs and the costs of compression.

Impurities can affect the economics of transport, because higher volumes of gas need to be transported, and liquefying a multiphase mixture is more costly. In addition to economic

considerations, removal of impurities may be required by existing health and safety legislation related to the compression, transport and injection of CO₂. Such economic and HSE requirements leave operators with inherent incentives to reduce or remove impurities to the extent necessary to operate economically and legally, making additional legislation unnecessary.

Figure S.1 gives an overview of those impurities that have to be removed or significantly reduced in a CO₂ stream for economically viable and safe operation of compression, transport and injection (top three yellow bars). It also highlights that when existing legislation already regulates impurities, or when there is an inherent incentive to remove or reduce contaminants at a particular stage in the CCS chain, there may be no need for additional regulation of the CO₂ stream at later stages.

Impurities (presence or conc. of) in captured stream	Inherent requirements for CCS chain steps	Regulation and allowable co-contaminants
CH ₄ SO ₂ H ₂ O N ₂ O ₂ Ar H ₂ S H ₂ S NO ₂ As CO Hg	Compression, including HSE	
SO ₂ Hg CO NO ₂ As H ₂ S	Transport, including HSE	
SO ₂ Hg CO NO ₂ As H ₂ S	Injection, including HSE	
? ? ? ? ? ? ? ?		"Leftover" regulation for impurities in the reservoir
Acceptable impurities and concentrations		

Figure S.1 A conceptual diagram illustrating different CCS chain step requirements for the reduction of impurities in the CO₂ stream. It also illustrates the uncertainties around the impacts of impurities in the geological storage reservoir.

There are more uncertainties and less economic incentives around the role of impurities once the CO₂ stream is injected underground (Figure S.1, lowest yellow bar). In addition, from a storage perspective, the issue of impurities need to be viewed in connection to the compression, transport, injection and health and safety requirements earlier in the CCS chain in order not to over-regulate the CO₂ stream. If quantitative limits on the CO₂ stream composition are deemed necessary in the future, such legal provisions need to strike a balance between the economics of the entire CCS chain and protecting people and the environment.

1. Introduction

Carbon dioxide capture and storage (CCS) is a technology with the potential to make deep cuts in greenhouse gas emissions to mitigate climate change (IPCC, 2005; IEA, 2010). Demonstrating the safe and permanent storage of CO₂ in geological storage complexes is an important factor for substantiating the global potential of CCS.

CO₂-rock interactions are investigated with models and laboratory experiments using pure CO₂. However, as a consequence of the capture process, the CO₂ stream is likely to contain impurities, which may alter the behaviour of the stream through the compression, transport and storage elements of the CCS chain. Research on the effects of these impurities on the integrity of geological storage formations is limited.

The 'other components' or 'impurities' in CO₂ streams¹ can include nitrogen (N₂), oxygen (O₂) and water (H₂O), but also air pollutants such as sulphur and nitrogen oxides (SO_x and NO_x), particulates, hydrochloric acid (HCl), hydrogen fluoride (HF), mercury, other metals and trace organic and inorganic contaminants (IPCC, 2005). In addition small amounts of chemical solvents used in post-combustion capture may be present in the CO₂ stream (Visser et al., 2008). The removal of certain contaminants may be required for health, safety and environmental protection reasons, but also to ensure the effective transport and storage of the CO₂ stream.

However, without overlooking the importance of health, safety and the protection of the environment, the extent to which impurities must be removed from the CO₂ stream is also an economic issue. Reaching higher levels of CO₂ purity will involve a number of incremental gas treatment processes each incurring capital and operation costs, potentially increasing the energy penalty and reducing CO₂ avoidance. It is important that if deemed necessary, legal provisions regulating the maximum levels of impurities permitted in a captured CO₂ stream strikes a balance between the economics of the entire CCS chain and protecting people and the environment. In order to do this, policymakers need access to reliable scientific research, review existing regulation and consult a range of stakeholders.

This report aims first and foremost to provide an overview of the current understanding regarding impurities in captured CO₂ streams and the transport and storage impacts of those impurities. Second, the report will explore possible EU regulation on the topic. The report also presents a detailed literature review of the likely CO₂ stream composition from capture, requirements for transport, and interactions with storage reservoirs.

The implications of CO₂ stream composition on compression requirements, pipeline design and operation has been thoroughly researched and resulted in a number of recommendations, either for devising European regulations or to be used as standalone best practice documents (Visser et al., 2008; DNV, 2010). These results are discussed in section 2.

There is also a growing body of research focusing on gas-rock interactions with the presence of impurities in geological storage formations (Gunter et al., 2000; Knauss et al.,

¹ The term impurities will be used in this report.

2005: Xu et al., 2007; Jacquemet et al., 2009; Koenen, et al., 2010). This literature is reviewed and discussed in section 3 of this report.

It is currently unclear to what extent the information on storage integrity and impurities can be utilised to support competent authorities in determining an acceptable CO₂ composition from capture projects. At present, European regulation does not specify in quantitative terms a legal composition of a captured CO₂ stream, and it is unclear whether this will be altered in the future. Some of the impurities may be self-regulated by industry, as their presence affects the economics of compression or transport. However, for other substances, the lack of an agreed specification could lead to uncertainties for industrial stakeholders of CCS projects (Birat, 2010) and possibly for storage integrity concerns. Conclusions and recommendations for further research and other questions are discussed in section 4.

2. Impurities in capture, compression and transport

2.1. *Impurities from capture installations*

The nature and quantity of impurities present in a stream of captured CO₂ is dependent on the fuels used, the capture process (the type of solvent used) and any gas treatment steps either prior or subsequent to CO₂ capture (Visser et al., 2008). The composition of the resultant CO₂ stream is contingent on the associated capture processes, and distinct differences in CO₂ stream composition exist between the three main categories of post-combustion, pre-combustion or oxyfuel capture (Anheden et al., 2004). CO₂ capture from industrial process, such as blast furnaces or cement kilns present another level of complexity in the stream composition, as the potential suitability of different types of capture technologies, either based on chemical adsorption or physisorption is much less understood.

The composition of the captured CO₂ stream resulting from the post-combustion capture process is understood to contain fewer impurities as existing regulations, specifically the Industrial Emissions Directive², which requires sets Emission Limit values (ELVs) on sulphur dioxide (SO₂), nitrogen oxides (NO_x) and dust particles. The removal of these pollutants will precede the removal of CO₂ from the remaining flue gas passing. The CO₂ stream after the CO₂ capture process will contain small (up to 0.3% by volume each) amounts of nitrogen, oxygen, argon, water and, in some cases, very small amounts of ash, trace metals, SO₂ and NO_x (< 0.01% by volume each; ICF International, 2010). Furthermore, elevated levels of NO_x (50ppm) has been proven to cause significant degradation of the monoethalineamine (MEA)³, the byproducts of which (such as alkanolamines, anionic heat stable salts and ammonia) will consequently reduce the efficiency of CO₂ removal (Pederson et al., 2010).

During pre-combustion capture, coal or biomass is gasified, resulting in a syngas composed of H₂ and CO, the latter which is converted to CO₂ through a water-gas shift process with the remaining hydrogen rich fuel used in a number of applications such as boilers, furnaces and gas turbines (Visser et al., 2007). The water gas shift process does result in a concentrated stream of CO₂ (>98%), however small amounts of hydrogen (H₂; 1.5%) and hydrogen sulphide (H₂S; ≈ 0.2%) may be present (ICF International, 2010). However, because gasification takes place in a reducing environment, no SO₂ or NO_x will be present in the captured CO₂ stream (Visser et al., 2007).

Dependent on feed composition and feed pressure, the purity of a CO₂ stream from an oxyfuel pulverized coal or natural gas installation can vary. The flue gas from an oxyfueled pulverized coal power plant contains approximately 63% CO₂, and thus needs to be purified (Pipitone & Bolland, 2009). The type and amounts of impurities present in the final CO₂ stream depends on the method of CO₂ purification. Incidental substances such as SO_x, NO_x, and Mercury (Hg), and significant amounts of nitrogen, argon and oxygen may be present in the captured CO₂ stream (ICF International, 2010). Oxyfuel combustion processes will also be regulated by the EU Large Combustion Plants Directive, which replaces restrictions on SO₂, NO_x and dust particles.

² Directive 2010/75/EU).

³ A common chemical solvent used in post-combustion capture systems.

In addition to CO₂ streams stemming from power generation processes, CO₂ capture processes may also be applied to industrial processes such as blast furnaces, cement kilns and certain parts of oil refineries. Capture techniques for these applications are currently in development, such as the oxyfuel cement kiln (IEA GHG, 2008) and the oxygen blast furnace with CO₂ capture. Even though at present it is still unclear which capture options are most suitable for various industrial processes, it is understood that high purity captured CO₂ streams can be achieved, albeit with additional energy and cost. Table 2.1 below provides an illustrative calculation of the compositions of CO₂ streams resulting from a range of capture processes from various installations.

Table 2.1 *Calculated compositions of CO₂ streams from a range of capture processes (adapted from ICF International, 2010)⁴*

Component	Post combustion capture at subcritical pulverized coal plant (MEA)	Pre-combustion capture at coal IGCC plant (Selexol)	Oxy-fuel combustion at supercritical pulverised coal plant (Stack gas)	Cement plant (MEA)	Refinery stack (MEA)
Argon, Ar	22 ppmv	178 ppmv	3.7%	11 ppmv	38 ppmv
Arsenic, As	0.0055 ppmv	0.0033 ppmv	0.0085 ppmv	0.0029 ppmv	
Methane, CH ₄		112 ppmv		0.026 ppmv	
Chlorine, Cl	0.85 ppmv	17.5 ppmv	0.14%	0.41 ppmv	0.4 ppmv
Carbon monoxide, CO		0.13%		1.2 ppmv	
Carbon dioxide, CO ₂	99.7%	98.1%	88.4%	99.8%	99.6%
Carbonyl sulphide, COS		1.7 ppmv			
Hydrogen, H ₂		1.5%			
Water, H ₂ O	640 ppmv	376 ppmv	640 ppmv	640 ppmv	640 ppmv
Hydrogen sulphide H ₂ S		0.17%			7.9 ppmv
Mercury, Hg	0.00069 ppmv	0.00068 ppmv	0.0035 ppmv	0.00073 ppmv	
Nitrogen, as N ₂	0.18%	195 ppmv	2.8%	893 ppmv	0.29%
Ammonia, NH ₃		38 ppmv			
Nitric oxides as NO ₂	1.5 ppmv	11 ppmv	721 ppmv	0.86 ppmv	2.5 ppmv
Oxygen, O ₂	61 ppmv		3.6%	35 ppmv	121 ppmv
Selenium, Se	0.017 ppmv	0.01 ppmv	0.026 ppmv	0.0088 ppmv	
Sulphur oxides, as SO ₂	0.84 ppmv		1.3%	0.097 ppmv	1.3 ppmv

2.2. Potential impacts of impurities on compression and transport

The presence of non-condensable gases in the CO₂ stream increases the costs of compression and transport, while effectively reducing the storage capacity of geological formations (Yan et al., 2009). In addition, water and compounds such as hydrogen sulphide (H₂S) lead to corrosion of pipelines. Note that such processes have safety impacts but also affect the economics of compression and the transport process of a CCS project.

⁴ Note that these concentrations refer to specific examples of modeled capture setups, and cannot be generalised to represent the different categories of capture processes in general.

2.2.1. Compression

Non-condensable gases in the CO₂ stream may include hydrogen (H₂), argon (Ar), nitrogen (N₂), oxygen (O₂) and methane (CH₄). These gases have a weak tendency to mix with other gases and dilute the CO₂, and complicate the phase change during compression. This means that higher pressures will be needed to avoid 2-phase flow (Visser et al., 2007; Anheden et al., 2004), leading to potentially substantial additional energy requirements. For example, 5% of H₂ present in the CO₂ stream composition represents an increase of 25% compression work (Austegard & Barrio, 2006). In the oxyfuel case in Table 2.1 above, approximately 8% of the stream is composed of Ar, O₂ and N₂, and this may require further gas conditioning prior to compression and transport.

2.2.2. Transport

Aside of the compression stage, non-condensable gases can also have an impact on CO₂ pipeline transport, both in terms of the volume available for CO₂, and the necessity for an increase in pressure to keep the CO₂ in dense phase⁵. For example, the volumetric capacity of a pipeline is decreased by 27% for a stream of CO₂ that contains 10% of hydrogen (Mohitpour, 2003). Yan et al., (2009), have evaluated the influence that various purities of CO₂ stemming from an oxyfuel capture installation may have on the costs of CO₂ transportation. Three purification scenarios were simulated resulting in CO₂ stream purities of 87%, 96% and 99%, with the remaining stream composition comprising of various amounts N₂, Ar and O₂. Although the lower purity CO₂ stream results in reduced capture costs (Δ2€/t CO₂ 87-99%), the overall chain costs for the 87% purity stream were €2.5/t CO₂ higher than the 99% purity stream. The augmented costs of transporting the 87% CO₂ stream are affiliated to the additional 23% compression work to avoid 2-phase flow, and an incremental increase of approximately 100mm in pipeline diameter to compensate for lost volume due to the impurities. According to Yan et al., the 96% purity stream is understood represent the optimum cost balance in the CCS chain.

Impurities present in a CO₂ stream can also lead to corrosion and hydrate formation within a pipeline. Free water present in the CO₂ stream can react with small amounts of CO₂, H₂S and CH₄ and result in the formation of hydrates. Hydrates are solid crystalline compounds that can block the pipeline and potentially damage equipment (Carroll, 2003). Within carbon manganese steel pipes, (C-Mn) free water combined with the high CO₂ partial pressure can cause extreme corrosion rates, primarily due to the formation of carbonic acid. Sulphuric acids can also form if the stream is co-contaminated with H₂S or SO₂. Under laboratory conditions, Choi et al., (2010) recorded dramatically increased corrosion rates of carbon steel from 0.38 mm/yr to 5.6 mm/yr, after the addition of 1% SO₂ to water-saturated CO₂. The rate was increased again to 7 mm/yr with the addition of O₂. However, through sufficient dewatering of the CO₂ prior to transportation, corrosion can be almost entirely avoided⁶.

⁵ The most efficient for transportation.

⁶ Corrosion rates are in the order of mm/y when free water is present, and can be reduced to the order of μm/y when dry CO₂ is transported (Seiersten, 2001).

Table 2.2 Descriptions and potential impacts of CO₂ and possible impurities for health and safety, compression and transport issues (adapted from DNV, 2010)

Component	Health and safety	Pipeline capacity	Water solubility	Hydrate formation	Materials	Fatigue	Fracture	Corrosion	Operations	Comment
CO ₂	•	•	•	•	•	•	•	•	•	Non-flammable, colourless, no odour; low toxicity, heavier than air in the gaseous state
H ₂ O				•	•	•	•	•	•	Non-toxic; condensable; forms acids with CO ₂ , NO _x and SO _x , which have a corrosive impact on transport infrastructure
N ₂		•	•							Non-toxic; stable
O ₂		•	•					•		Non-toxic
H ₂ S	•	•			•	•	•	•		Flammable, strong odour, extremely toxic at low concentrations
H ₂		•	•				•			Flammable, non-condensable at pipeline operating condition; impact on transport infrastructure through embrittlement
SO ₂	•		•					•		Non-flammable, strong odour, toxic; forms sulphuric acid with water
NO ₂	•		•					•		Non-flammable, toxic; forms nitric acid with water
CO	•		•							Flammable, toxic
CH ₄ +		•	•						•	Odourless, flammable
Amines	•									Potential occupational hazard with corrosive impact
Glycol	•							•		Potential occupational hazard

2.3. Regulation of captured CO₂ stream composition

A solitary piece of European regulation currently comments on the requirements for the composition of a captured stream of CO₂. The European Union Directive on the geological storage of CO₂⁷, defines the CO₂ stream as ‘a flow of substances that results from CO₂ capture processes.’ The Directive also loosely defines the required stream composition that can be legally transported. Article 12(1) states in part:

‘A CO₂ stream shall consist overwhelmingly of carbon dioxide. To this end, no waste or other matter may be added for the purpose of disposing of that waste or other matter.’

Although clearly prohibiting the co-disposal of waste gases in a CO₂ stream, the Directive does not set absolute quantitative restrictions on the substances that compose the CO₂ stream, but uses qualitative criteria. The Directive does recognize that the CO₂ stream may contain incidental associated substances from the capture process, or substances used for monitoring and verification purposes, however all incidental or added substances must be below levels that would:

⁷ Directive 2009/13/EC – henceforth referred to as ‘the Directive’.

- a) *'adversely affect the integrity of the storage site or the relevant transport infrastructure;*
- b) *pose a significant risk to the environment or human health; or*
- c) *breach the requirements of applicable Community legislation.'*

To meet the above criteria, operators are required to carry out a risk assessment in respect of the stream composition and maintain a register of the quantity, properties and composition of streams injected (UCL, 2010). However with reference to point "a)" above, questions can be raised regarding how various compositions of CO₂ streams may affect the integrity of the storage site. It is currently unclear whether sufficient information exists to allow a well-informed decision about the level of impurities that can be injected into a storage site.

In legal terms, the use of qualitative criteria for a gaseous stream composition (which can be quite easily quantified) seems inappropriate, as the term 'overwhelmingly' used in Article 12(1) could be interpreted differently between operators. The term 'overwhelmingly' was initially utilized in the London Protocol, the first international marine-environment protection instrument that permits the offshore geological storage of CO₂ (Holwerda, 2011). In the Directive, the formula that CO₂ streams "consist overwhelmingly of carbon dioxide" was chosen as it allows for a case-by-case assessment of the levels of impurities, recognizing the natural variation in storage site characteristics and different transport constructions⁸.

Brockett (2009) informs that during the drafting of the Directive, the European Parliament's Environment Committee had proposed an amendment to the Directive, calling for a CO₂ concentration of ≥95% and above, and the elimination of H₂S and SO₂. This amendment was not adopted, on the basis that certain applications of CCS, particularly for the cement and steel industry⁹, may have considerable problems reaching such levels of CO₂ purity. Furthermore the complete removal of H₂S and SO₂ is potentially impossible.

The Commission has produced guidance on the practical applications of the qualitative criteria outlined in Article 12 of the Directive¹⁰. It is also expected that documents specifying the Best Available Techniques (BAT) will be developed for capture installations, and these documents would include CO₂ compositions (Brockett, 2009). The Directive, including the provisions of Article 12, will also be reviewed and modified if required in 2015.

2.4. Recommendations on the CO₂ stream composition

The European DYNAMIS project has assessed the requirements for CO₂ stream composition in terms of health and safety, and the durability of the transport network (Visser et al., 2007). With exemption of the CO₂ purity levels suitable for use in enhanced oil recovery operations, the impact of impurities on geological storage formation or the behaviour of the CO₂ plume was not covered. Nevertheless, in order to best inform policymakers on any necessary regulatory requirements on CO₂ purity for storage, an

⁸ *International Marine Regulation*, supra note 22, at 4506.

⁹ CCS applications in many industrial sectors, particularly the steel and cement sectors, are currently in the early stages of development. There is not sufficient information on the CO₂ reduction potential of such industrial process with CCS, and the costs associated with such applications are uncertain.

¹⁰ See - http://ec.europa.eu/clima/policies/lowcarbon/docs/gd2_en.pdf.

assessment of the whole CCS chain is necessary. The compositional demands of CO₂ for storage cannot be viewed in isolation from compression, transport and safety.

In terms of transporting a CO₂ stream in a safe way, substantial amounts of impurities present in the stream may influence the impacts of a pipeline leakage or rupture (IPCC, 2005). Furthermore, in addition to the dangers of a sudden high concentrated release of CO₂, compounds such as CO, SO₂ and H₂S which may also be present in the stream are toxic by nature. Visser et al. (2007) used a set of calculations based on European Short Term Exposure Limits (STEL) to determine the maximum recommended concentrations of H₂S, CO, SO₂ and NO₂ in pipelines. A precautionary multiplication safety factor of 5 was used, justified on the basis of uncertainties in the diffusion patterns and possible synergetic effects between the various purities.

The DYNAMIS project also sets recommended technical limits for impurities in the CO₂ stream. The formation of hydrates and subsequent corrosion of the pipeline, and the reduced pipeline capacity for CO₂ are raised as the main negative impacts brought about by the presence of impurities in the CO₂ stream. The recommended composition of a CO₂ stream for efficient and safe transportation from the DYNAMIS project is presented in Table 2.4.

Table 2.3 *Recommendations for CO₂ composition for safe transport (adapted from Visser et al., 2007)*

Component	concentration	Limitation
H ₂ O	500 ppm	Technical: below solubility limit of H ₂ O in CO ₂
H ₂ S	200 ppm	Health and safety considerations
CO	2000 ppm	Health and safety consideration
O ₂	< 4%	Precautionary due to unknown effects underground
CH ₄	< 4%	Methane affects the solubility of water
N ₂	< 4%	Non-condensable gas effect storage/transport capacity
Ar	< 4%	Non-condensable gas effect storage/transport capacity
H ₂	< 4%	Non-condensable gas effect storage/transport capacity/ economic reasons (energy content)
SO _x	100 ppm	Health and safety considerations
NO _x	100 ppm	Health and safety considerations
CO ₂	>95.5%	Balanced with other compounds

All recommendations made by Visser et al., (2007) are based on criteria focusing on health and safety, and the efficient transport of CO₂ in pipelines. To date, there have been no specific recommendations presented to European regulators based on foreseen impacts of impurities on long-term geological storage formations. There have been however, a number of scientific investigations completed to attempt to explore, and if possible quantify such impacts. The results of such investigations are documented in section 3 of this report.

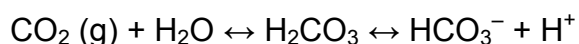
3. Current understanding on the impact of impurities on CO₂ gas-rock interactions

According to the IPCC (2005), there are three main geochemical trapping processes that can occur at different times during the storage process: solubility trapping, ionic trapping and mineral trapping. Before moving directly to the impacts of impurities on gas-rock interaction, it's important to cover the basic aspects of geochemical trapping of CO₂.

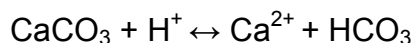
3.1. *Geochemical trapping of CO₂*

Saline reservoirs in sedimentary basins can provide suitable geological formations for the safe storage of supercritical CO₂ (; IPCC, 2005; Gaus, 2010). The salinity of water present in such formations is understood to range from 5,000 to over 350,000 mg/L, and the majority are not considered potable given the presence of dissolved toxic materials, organic and/or inorganic components naturally occurring radioactive material as well as the high salinity itself (Kharaka & Hanor, 2007). The conditions found with such formations are the result of various hydrogeochemical processes including mixing, dissolution/precipitation of minerals, bacterial activity and interactions with organic material. The injection of CO₂ into subsurface geological formations may induce a number additional geochemical processes, potentially altering the chemical reactivity of the system (Gaus, 2010).

Once injection has started, supercritical CO₂ and any other potential impurities will dissolve and interactions will occur between the injected stream and the well materials (Gaus, 2010). Although CO₂ alone is not reactive, once in contact with brine, carbonic acid (H₂CO₃) is formed, which dissociates in the brine. The formation of this acid reduces the pH of the brine, meaning that the brine becomes corrosive to well materials, rocks and pipelines (Gaus, 2010).



This reaction is then buffered through interactions with carbonate minerals present in the geological formation, the acidity of the brine will still be sufficient to react with other carbonate and silicate minerals (Gaus, 2010).



Dependent on the geology of the storage formation, the acid may react with calcium, magnesium and iron carbonates to form bicarbonate ions (IPCC, 2005). The trapping of CO₂ by dissolving it into the formation water is known as solubility trapping. Solubility trapping is understood to benefit the security of CO₂ storage because as CO₂ is dissolved, it no longer exists as a separate phase, losing the buoyant forces that can drive the CO₂ plume upwards. The process of solubility trapping is thought to be the dominant geochemical process during the first tens to hundreds of years after injection (IPCC, 2005).

The incorporation of the injected CO₂ into minerals due to chemical precipitation is termed mineral trapping, and this can be a rapid (a number of days) or a slow process (thousands of years) depending on the type of mineral (IPCC, 2005), but also on the in-situ

temperature and pressure (Kaszuba et al., 2005). Examples of minerals that may precipitate in geological formations which contain clays and feldspars include, Dawsonite (from feldspar albite), siderite and dolomite (from the clay chlorite) (Gaus, 2010).

3.2. Dissolved gas-rock interactions and injectivity

Injectivity is the rate at which CO₂ can be injected into a given reservoir and the ability of the subsequent CO₂ plume to migrate away from the injection well (Kaldi and Gibson-Poole, 2008). There is a wide range of factors that determine the injectivity both anthropogenic, for example; wellbore design, injection rate and injection pressure, and natural; viscosity of in-situ fluid, stratigraphic architecture and the formation porosity and permeability characteristics (Soloman and Flach, 2010).

Permeability is one of the main controlling parameters of CO₂ storage in geological formations (van der Meer, 1995; Law and Bachu, 1996; Ennis-King and Paterson, 2001). If the formation permeability is insufficient, the low permeability will reduce injectivity with consequential negative impacts on overall CO₂ storage capacity. This could indicate that a potential storage location is not commercially viable. Although permeability can be tested prior to CO₂ injection, once injection has commenced, injectivity losses could arise from geochemical reactions between the CO₂ saturated formation water, but also from the presence of gaseous impurities in the CO₂ stream (Soloman and Flach, 2010).

Geochemical gas-rock interactions such as solubility trapping and mineralization are important for the permanent storage of the injected CO₂. However, although CO₂-saturated brine can initially cause increased porosity and permeability, insoluble reaction products will be precipitated downstream, potentially plugging pore throats and consequently reducing permeability (Soloman and Flach, 2010). According to Gaus et al., (2010), drastic permeability reductions due to minute amounts of dissolved minerals affecting the geometry of hydraulic capillaries are possible. To calculate the porosity change brought about by chemical precipitation, the balance between the space created by the dissolved mass and the space occupied by the precipitated minerals needs to be assessed (Gaus et al., 2010).

3.3. The impact of impurities on geological formations

3.3.1. Storage capacity

Impurities in a CO₂ stream, such as non-condensable gases, reduce the density of the gas stream. The lower density leads to a consequential drop in the total storage capacity of a storage reservoir. A less pure CO₂ stream would mean that storage locations would be expended at a faster rate, meaning that higher costs would be incurred both due to the physical pore space occupied, but also through the costs of more frequent re-mobilising of injection equipment, re-installation of sub-surface templates and conducting additional well characterizations. Overall, it is of course desirable to utilize geological storage space as efficiently as possible.

Yan et al., (2009) conducted a techno-economic assessment developing three CO₂ purification scenarios (87%, 96% and 99% CO₂), stemming from an oxyfuel combustion installation. As well as compression and transport costs, the effects of non-condensable gases (N₂, Ar and O₂) on storage capacity was also investigated. Figure 3.1 depicts the results of a simulation, highlighting the density changes of the three CO₂ streams as they are injected to various depths. The graph on the left is a simulation of injecting a stream of

CO₂ to a depth of 2000m, whereas on the rate the depth is 1000m, with corresponding values of the y-axis.

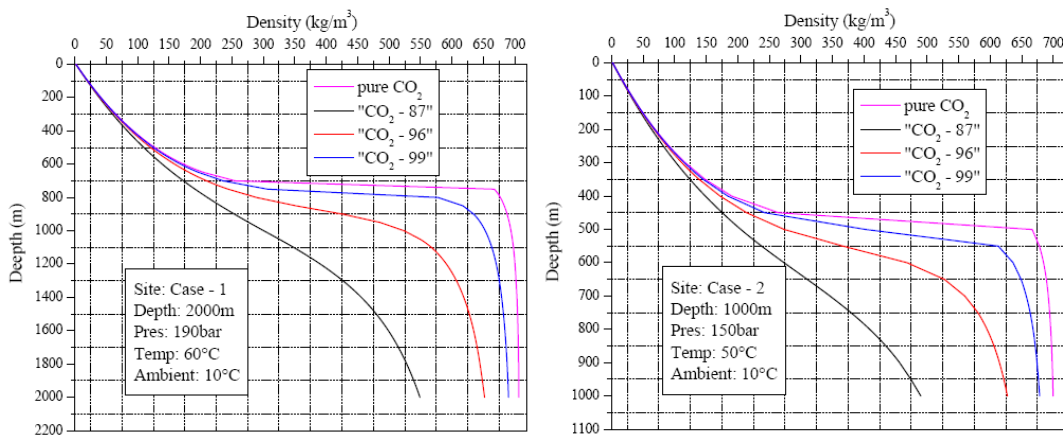


Figure 3.1 The density of CO₂ streams with different purities of CO₂ at different depths, pressures and temperatures (Yan et al., 2009)

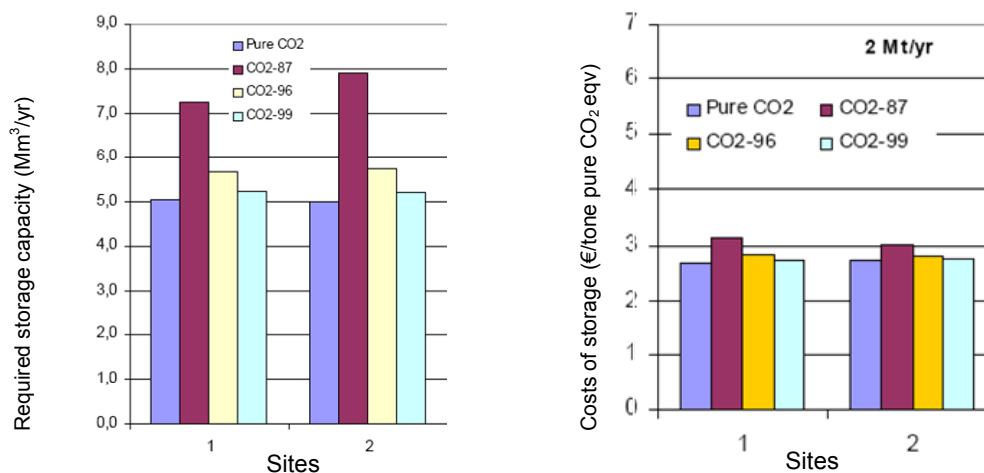


Figure 3.2 The required storage capacity (left) and associated costs (right) of CO₂ streams with different purities (Yan et al., 2009)

In the assessment, three storage scenarios with altered depths and geothermal conditions were presented, of which two are presented in Figure 3.1. It is clear that according to Yan et al., (2009), the density of the CO₂ stream at both depths of 2000m and 1000m is directly affected by the increased presence of non-condensable gases. Given that additional amounts of a low purity CO₂ stream must be injected to achieve a specified level of CO₂ storage, the costs of storage are approximately €0.3/tonne of pure CO₂ equivalent higher than in the 99% CO₂ stream purity (Figure 3.2). The costs include the CAPEX, OPEX and annual amortisation, although it is unclear if the injection takes place on or offshore.

The presence of impurities may also have an impact on the solubility and subsequently the geochemical reactions. According to Jacquemet et al., (2009), if such reactions lead to change in porosity relative to the porosity prior to injection, the space volume of the storage would be modified as well. Reductions in the interfacial forces between the fluid and the rock brought about by impurities could consequently reduce the sealing capacity and the storage efficiency (Jaquemet et al., 2009).

3.3.2. Induced porosity and permeability changes

As mentioned above, geochemical reaction between the CO₂ gas stream, in-situ brine and minerals in the storage formation can lead to problems during the operational phase, such as reduced permeability and increased pore pressures (Anheden et al., 2004). The blocking of fractures and pore spaces by mineral precipitation is a common feature in many geological settings (Emmanuel & Berkowitz, 2007). It is the growth in secondary minerals in the brine resulting from the precipitation process that are able to alter the geometry of the hydraulic capillaries. Impurities in the injected gas stream can have direct impacts on the intensity and/or nature of the gas-rock interactions, rapidly reducing the pH of the brine and potentially increasing dissolution of the host rock and precipitation of secondary minerals over time scales ranging from months to thousands of years (Anheden et al., 2004). However, it must also be raised that, depending on numerous in-situ conditions, modifications in permeability brought about by mineral reactions may actually aid the migration of CO₂ through the injection zone (ICF International, 2010).

If pure CO₂ is injected into a storage formation, once in contact with water a relatively weak carbonic acid (H₂CO₃) will be formed. However, if other compounds are co-injected with the supercritical CO₂, much stronger acids may develop. A number of the most potentially important acids and their relative acidity compared to carbonic acid have been calculated by ICF International (2010), and are presented in Table 3.1. The equilibrium constants used to develop the relative acidities have been calculated at 25°C and at atmospheric pressure, and thus may vary given geological conditions.

Table 3.1 *A range of acids and their relatively acidity compared to carbonic acid (adapted from ICF International, 2010)*

Acid	Formula	Relative acidity
Hydrochloric acid	HCl	$2.3 \times (10)^{14}$
Sulphurous acid	H ₂ SO ₃	$3.5 \times (10)^4$
Sulphuric acid	H ₂ SO ₄	$2.8 \times (10)^4$
Carbonic acid	H ₂ CO ₃	$1.0 \times (10)^0$
Nitrous acid	HNO ₂	$1.0 \times (10)^3$

There are a limited number of studies that attempt to quantify the effects that stronger acids may have on the precipitation of minerals in geological formations. Knauss et al., (2005), coupled a chemical model with a simplified flow in a one dimensional (1D) simulation using the reactive transport code CRUNCH (Steefel, 2001; Steefel and MacQuarrie, 1996), in order to simulate the injection of CO₂ into a heterogeneous rock formation, calculating the mineralogical changes over the flow path of 1km. In addition to modelling the impact of the CO₂, the simulation, based on an injection period of 5 years, was run with the addition of small amounts of H₂S and SO₂. The amount of SO₂ added was sufficient to reduce the pH to 1, given the experimental reservoir conditions. The results suggest that the co-injection of H₂S should not adversely impact injection rates compared to pure CO₂. However if sulphur is able to oxidise to sulphate, which is highly probable given the abundance of oxidants such as water, a sulphuric acid will be formed. This strong acid reduced the pH in the testing domain, and consequently led to significant precipitation of calcite, dawsonite and anhydrite.

Further 1D reactive transport simulations made by Xu et al., (2007), using a different code called TOUGHREACT (Xu et al., 2004), focused on the injection of CO₂ in sandstone formations. The results of the simulation highlight the dominant geochemical impact of SO₂ compared to pure CO₂ or CO₂ and H₂S, supporting the findings of Knauss et al., (2005). In the case of co-injection of CO₂ and SO₂, 0.036 kg/s of SO₂¹¹ was injected with 1 kg/s of CO₂, approximately 2.4% of the total injected material by mass¹². The illustrative graph below displays the approximate pH values of the formation fluid 10 years after injection, also displaying the changes in pH in relation to the distance from the injection zone.

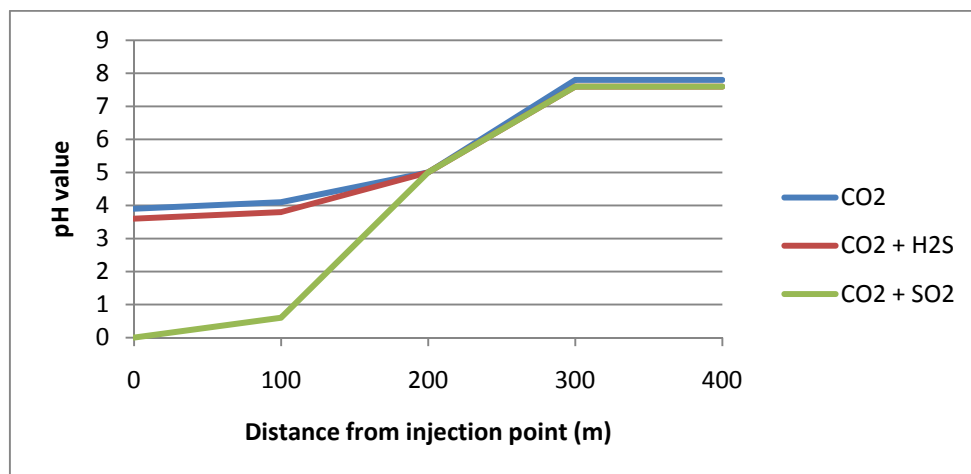
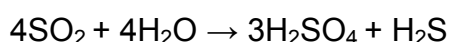


Figure 3.3 Changes in the pH of formation fluid after injection of pure CO₂, and combinations of CO₂, H₂S and SO₂ (graph modified from Knauss et al. 2005)

As can be seen from Figure 3.3, the combination of CO₂ and SO₂ leads to a dramatically reduced pH within the first 200 metres from the injection point. The combination of CO₂ and H₂S has a relatively small impact on pH compared to the base case. The acidic conditions are caused by the formation of sulphuric acid (Knauss et al., 2005):



The area of extreme acidity covers a radial distance of 100m from the point of injection. The lower pH due to the presence of SO₂ causes the rapid dissolution of calcite within the first 200m. The dissolution of calcite is followed by the precipitation of both minerals anhydrite and alunite, with the volume of precipitated minerals peaking at around 200m from the injection point. In terms of changes in porosity, the increased acidic conditions actually increased the porosity in the first 100m from the injection point. However, coinciding with the peak abundance of precipitated mineral, the porosity is lower than the base case at 200m 100 years after injection. Xu et al., (2007) suggest that small changes in porosity could result in a significant change in permeability, altering the flow pattern of the injected CO₂.

¹¹ This amount was chosen as it represents an average weight ratio of sulphur to carbon in a domestic coal fired power plant of 2.5:100 (Apps, 2006).

¹² A significant amount of H₂O was co-injected in addition to CO₂ in order to partly replicate the composition of the formation brine.

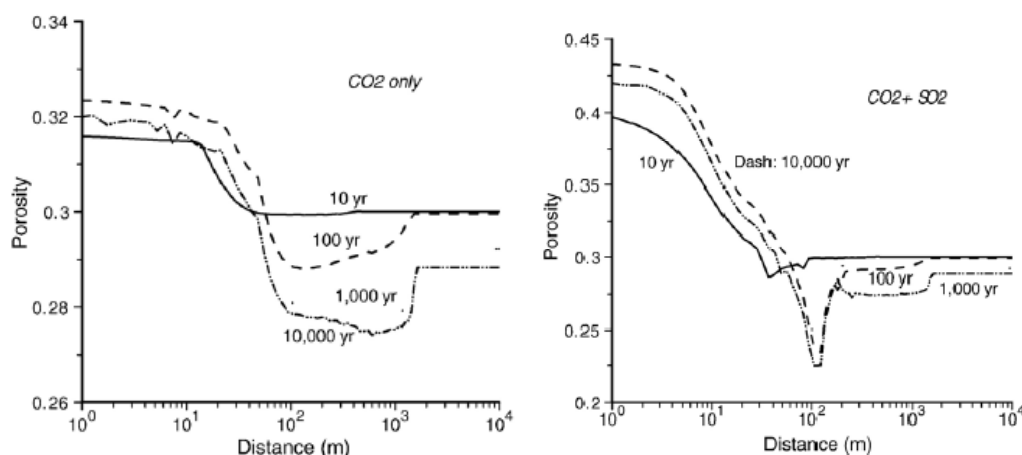


Figure 3.4 Modelled changes in porosity after injection of pure CO₂ (left) and a combination of CO₂ and SO₂

Figure 3.4 (Xu et al., 2007), compares the modelled porosity changes between the base case (CO₂ only) and the combination of CO₂ and SO₂. At between 100 to 200m from the injection point, both scenarios display reduced porosity for all timeframes. It is also clear that in the case of CO₂ and SO₂, the minimum porosity reached at approximately 200m is approximately 0.07 pp lower than the control at 100 years. The reduced porosity in both cases coincides with the maximum precipitation of alunite.

Koenen et al., (2010) have also investigated the impact of impurities on a sandstone formation. The investigation involved the use of the geochemical modelling tool PHREEQC, which simulated the injection of CO₂ streams into a formation with the mineralogical composition of a potential Dutch CO₂ storage field. Three CO₂ composition scenarios were modelled; a pure CO₂ base case, and scenarios based on the expected compositions of captured CO₂ streams from pre-combustion and oxy-fuel technologies. The composition of the CO₂ streams from the pre-combustion and oxyfuel technologies are shown in table 3.2.

Table 3.2 The composition of CO₂ streams used in the modelling study of Koenen et al. (2010)

Component	Pre-combustion (mole %)	Semi-purified oxyfuel (mole %)
CO ₂	99.64	98.0
H ₂ S	0.00014	-
CO	0.03	0.005
O ₂	0.0045	0.7
N ₂	0.077	0.7
Ar	-	0.6
H ₂	0.14	-
SO ₂	-	0.007
NO _x	-	0.01

The results of the model experiment in all cases display a drop in pH due to the formation of acids. In the base case, the pH drops from 6.1 to 4.6 due to the formation of carbonic acid. In the semi-purified oxyfuel case, the formation of sulphuric and nitric acids (through the reaction between SO₂, NO_x, H₂S and water) leads to a slightly lower pH of 4.5. Given

the small changes in pH, the difference in dissolution and consequently permeability is negligible. However, by increasing the concentrations of the impurities in the oxyfuel scenario tenfold, simulating an accumulation of impurities at the injection point, the pH drops to approximately 1.8. The acidification of the pore water led to a slightly higher calcite dissolution on the short term, though the porosity increase was negligible. Furthermore, the calcite dissolution also acts as a buffer, with the final pH raising again to 4.5. No precipitation of anhydrite (CaSO₄) was observed in this work, dissimilar to the results of Knauss et al., (2005) and Xu et al., (2007) covered above.

Of the main literature sources reviewed, there are a number of consistent findings. Dependent on reservoir geology, the co-injection of SO₂ with the CO₂ reduces the pH of the formation water to approximately 1, due to the formation of sulphuric acid. H₂S is seen to have a lesser impact on the creation of acidic conditions. With the co-injection of SO₂, a highly acidified zone forms within a radial distance of 200m from the injection point. In this acidic zone, rapid mineral dissolution of carbonate and silicate minerals may actually increase the porosity. At the edge of the injection zone (between 150-200m), the increased pH results in the precipitation of various secondary minerals, for example dawsonite, alunite, siderite and ankerite (dependent on the lithology of the formation). After approximately 100 years the level of precipitation is understood to reduce the porosity and potentially the permeability of the formation. A reduction in permeability could modify fluid flow, however this would not impact injectivity as the operational phase of the storage process would be complete.

4. Conclusion

Within the EU Directive on the geological storage of CO₂, the absence of quantitative requirements for the composition of a CO₂ can be attributed to a number of factors. Firstly, given that CCS is novel technology both in the power sector and industry, there is uncertainty regarding the composition of CO₂ streams that could be captured from currently unproven technologies. An eventual deployment of CCS may see a wide range of capture techniques being deployed, all resulting in heterogeneous gas stream compositions. In light of this, having a single purity requirement may incur significant economic penalties for certain technologies and hinder the development of CCS as a potential CO₂ abatement option both in the energy and industrial sector.

From another point of view, the implementation of multiple 'CO₂ purity classes' may obstruct the development of CO₂ transport networks and complicate the establishment of a 'captured carbon' commodity market (Paxton, 2011). It has been highlighted that if CCS is to significantly contribute to reducing CO₂ emissions in Europe, networks of pipelines carrying CO₂ from numerous installations will be required. Having some sort of minimum threshold or compositional ranges for impurities may then be required to facilitate the co-utilisation of transport and storage infrastructures. Furthermore, if Member States were to adopt divergent quantitative limits for impurities, this could impede cross-border trade of captured CO₂ and hamper the free movement of goods within Europe (Holwerda, 2011).

The second factor is the lack of understanding on the affects of impurities on various parts of the entire CCS chain. This is complicated by the multitude of possible interactions between the CO₂ stream and transport/storage equipment. Table 2.2 provides a useful overview of the potential impacts that impurities may have on CO₂ transport infrastructures, and also highlights health and safety concerns. However publicly available material covering the impacts of impurities on the storage of CO₂ appears limited.

The objective of this report has been to review current scientific literature regarding the implications of storing a CO₂ stream with the presence of impurities. Prior to the completion of this report, no publicly available review of this subject could be found.

4.1. Main findings

Capture processes rarely result in pure CO₂, which poses challenges for compression, transport and storage. Two possible impacts of impurities on CO₂ storage processes were identified:

- Non-condensable gases affecting storage capacity
- The formation of strong acids, which reduces the pH of the brine solution with possible affects on porosity and permeability

The presence of non-condensable gases such as H₂, Ar, N₂ and O₂ can significantly reduce the density of the CO₂ stream. In particular, captured CO₂ streams from oxyfuel capture technologies can contain relatively large percentages of non-condensable gases, up to approximately 10% depending on the conditioning processes deployed at the installation. For transport purposes, the incidence of non-condensable gases in the CO₂ stream would result in the requirement for additional compression work. Specifically for storage, the low density of the injected CO₂ stream would lead to the inefficient

utilisation of pore space, reducing the amount of CO₂ that can be stored at a particular storage location. Therefore, there are economic reasons to partially remove certain contaminants in the CO₂ stream.

The presence of certain impurities such as H₂S, NO_x and SO_x, can increase the range and rate of geochemical reactions through the formation of hydrochloric, sulphuric and nitrous acids which reduce the pH of the formation water. From the scientific literature reviewed in this report, the presence of SO₂ and the formation of sulphuric acid, has the most potent effect at reducing the pH of the in-situ water (see Figure 3.3). The drop in pH results in an acidified radial zone of approximately 200m around the injection point, whereby the dissolution of carbonate material takes place. At the periphery of the acidified zone where the pH rises, the precipitation of secondary minerals is observed. The precipitation of such minerals is higher in the cases where SO₂ is co-injected with CO₂. After approximately 100 years, sufficient precipitation of secondary minerals may take place that can induce changes in porosity and permeability. However, no current scientific literature points towards the co-injection of SO₂ or H₂S leading to short-term (<10 years) porosity reduction that could hinder injectivity.

4.2. Implications for EU CO₂ storage legislation

This report has established that impurities in the CO₂ stream can have impacts on storage efficiency and the porosity of storage reservoirs, however it still remains a difficult task to recommend quantitative limits on the presence of impurities for the purposes of safe and efficient storage. Regarding storage capacity, work completed by Yan et al. (2009) supports previous recommendations by Visser et al. (2007), that the amount of non-condensable gases in the CO₂ stream should not exceed 4% by volume. This figure is understood to reflect an optimum balance between gas conditioning costs and the costs of compression.

In terms of the potential for impurities to induce changes in the porosity and injectivity of a storage site, the formation of acids from compounds such as SO_x, NO_x and H₂S are well understood. Moreover, work by Knauss et al., (2005), Xu et al., (2007) and Koenen et al., (2010) provide initial quantitative insights into the magnitude of porosity changes given different amounts of impurities. Further research should be targeted towards understanding the kinetics of diffusion and chemical reactions to assess if impurities may accumulate at the injection zone, and whether impurities may alter the integrity of the caprock and well cement (Koenen et al., 2010).

Impurities (presence or conc. of) in captured stream	Inherent requirements for CCS chain steps		Regulation and allowable co-contaminants
CH ₄ SO ₂ H ₂ O N ₂ O ₂ Ar H ₂ S H ₂ S NO ₂ As CO Hg	Compression, including HSE		
SO ₂ Hg CO NO ₂ As H ₂ S	Transport, including HSE		
SO ₂ Hg CO NO ₂ As H ₂ S	Injection, including HSE		
? ? ? ? ? ? ? ?			"Leftover" regulation for impurities in the reservoir
Acceptable impurities and concentrations			

Figure 4.1 A conceptual diagram of possible CO₂ stream requirements at different stages of the CCS chain. It also illustrates the uncertainties around the impacts of impurities in the geological storage reservoir¹³.

As an extension of this report, it may be useful to question whether the development of storage specific legislation with regards to the possible impacts of impurities is necessary. As depicted in the conceptual diagram above, any CO₂ quality specifications for storage will be downstream from other purity requirements at different stages of the CCS chain such as compression and transport, as well as health and safety thresholds. One possibility is that upstream purity requirements render the development additional storage specific stream specifications as unnecessary.

In 2015, the European Commission shall review the Directive on the geological storage of CO₂, including an assessment of the provision on CO₂ stream acceptance and procedure referred to in Article 12 of the Directive. It is recommended that research on the behaviour of impurities in geological formations continues in order to provide further scientific findings to assist in the review process.

¹³ Please note that this is a illustrative diagram and the presence of certain impurities may not be accurate for certain capture processes.

5. References

- Anheden, M., Andersson, A., Bernstone, C., Eriksson, S., Yan, J., Liljemark, S., Wall, C., 2004. CO₂ quality requirement for a system with CO₂ capture, transport and storage. In: GHGT-7, Vancouver, September 5-9, 2004.
- Apps, J.A., 2006. A review of hazardous chemical species associated with CO₂ capture from coal-fired power plants and their potential fate during CO₂ geologic storage. Lawrence Berkeley National Laboratory Report LBNL-59731.
- Austegard, A. and M. Barrio (2006) Project Internal Memo DYNAMIS: Inert components, solubility of water in CO₂ and mixtures of CO₂ and CO₂ hydrates, October 2006.
- Birat, J.P., 2010. Sectoral assessment for CCS in the steel industry. United Nations Industrial Development Organisation.
- Brockett, S. 2009. The EU enabling legal framework for carbon capture and geological storage. *Energy Procedia* **1**, 4433-4441.
- Carroll, J.J. (2003) Problem is the result of industry's move to use higher pressures, *Gas Liquids Engineering Ltd., Pipeline & Gas Journal*, 230(6), 60-61, Calgary, Alberta, Canada.
- Choi, Y.S. and Nešić, S. (2010) Effect of impurities on the corrosion behavior of carbon steel in supercritical water environments. *Proceedings of Corrosion 2010*, March 14-18, San Antonio, TX.
- DNV, 2010. Design and operation of CO₂ pipelines - Recommended practice DNV-RP-J202. Det Norske Veritas, April 2010.
- Emmanuel, S. and B. Berkowitz (2007). Effects of pore-size controlled solubility on reactive transport in heterogeneous rock, *Geophysical Research Letters*, 34.
- Ennis-King, J. & Paterson, L. 2001. Reservoir engineering issues in the geological disposal of carbon dioxide, in Williams D. J., Durie, R.A, McMullan, P., Paulson, C.A.J. & Smith, A. Y. (eds), *Proceedings of the fifth international conference on greenhouse gas control technologies: GHGT5*, Collingwood, VIC, CSIRO Publishing, 290-295.
- Gaus, I., 2010. Role and impact of CO₂-rock interactions during CO₂ storage in sedimentary rocks. *International journal of greenhouse gas control* **4**, 73-89.
- Gaus, I., Audigane, P., André, L., Lions, J., Jacquement, N., Durst, P., Czernichowski-Lauriol, I. and Azaroual, M. 2010. Geochemical and solute transport modelling for CO₂ storage, what to expect from it? *International journal of greenhouse gas control* **2**, 605-625.
- Gunter, W.D., Perkins, E.H., Hutcheon, I., 2000. Aquifer disposal of acid gases: modelling of water-rock reactions for trapping of acid wastes. *Applied Geochemistry* **15**, 1085-1095.

Holwerda, M. (2011). Deploying carbon capture and storage “safely”: The scope for Member States of the EU to adopt more stringent CO₂ stream-purity criteria under EU law. *Climate Law*, 2, 37-61. DOI 10.3233/CL-2011-025.

ICF International, 2010. Implementation of Directive 2009/31/EC on the Geological Storage of Carbon Dioxide - Guidance document 2: Site characterization, CO₂ stream composition, monitoring and corrective procedures. Available at (21/10/2010): [http://ec.europa.eu/clima/policies/lowcarbon/docs/GD2-site%20characterisation-CO₂%20stream-monitoring-corrective%20measures-consultation.pdf](http://ec.europa.eu/clima/policies/lowcarbon/docs/GD2-site%20characterisation-CO2%20stream-monitoring-corrective%20measures-consultation.pdf)

IEA, 2004. Prospects for CO₂ capture and storage. International Energy Agency Publications. Paris, France.

IEA, 2008. CO₂ capture and storage. A key carbon abatement option. IEA: Paris, France.

IPCC, 2005. IPCC special report on carbon dioxide capture and storage. Prepared by Working Group III of the Intergovernmental Panel on Climate Change [Metz, B., O. Davidson, H. C. de Coninck, M. Loos, and L. A. Meyer (eds.)]. Cambridge University Press, Cambridge, United Kingdom and New York, NY, USA, 442 pp.

Jacquemet, N., Le Gallo, Y., Estublier, A., Lachet, V., von Dalwigk, I., Yanc, J., Azaroual, M., Audigane, P., 2009. CO₂ streams containing associated components — a review of the thermodynamic and geochemical properties and assessment of some reactive transport codes. *Proceedings of the 9th International Conference on Greenhouse Gas Control Technologies (GHGT-9)*, November 16-20, 2008, Washington, DC, USA, *Energy Procedia* 1 (1), 3739-3746.

Johnson, J.W., Nitao, J.J., Steefel, C.I., Knaus, K.G., 2001. Reactive transport modeling of geologic CO₂ sequestration in saline aquifers: The influence of intra-aquifer shales and the relative effectiveness of structural, solubility, and mineral trapping during prograde and retrograde sequestration, In *proceedings: First National Conference on Carbon Sequestration*. Washington, DC.

Kaldi, J.G., & Gibson-Poole, C.M. 2008. Storage capacity estimate, site selection and characterisation for CO₂ storage projects, CO₂CRC Report No. RPT08-1001, Canberra, ACT, Cooperative Research Centre for Greenhouse Gas Technologies (CO₂CRC).

Kaszuba, J.P., Janecky, D.R., Snow, M.G., 2005. Experimental evaluation of mixed fluid reactions between supercritical carbon dioxide and NaCl brine: relevance to the integrity of a geologic carbon repository. *Chemical Geology* 217, 277-293.

Kharaka, Y.K., Hanor, J.S., 2007. Deep fluids in the continents: I. Sedimentary basins. In: Drever, J.I. (Ed.), *Surface and Ground Water, Weathering and Soils, Treatise on Geochemistry*, vol. 5. Elsevier, pp. 1-48.

Knauss, K., Johnson, J.W., Steefel, C.I., 2005. Evaluation of the impact of CO₂, cocontaminant gas, aqueous fluid and reservoir-rock interactions on the geologic sequestration of CO₂. *Chemical Geology* 217, 339-350.

Koenen, K., Tambach, T. and Neele, F. 2010. Geochemical effects of impurities in CO₂ on a sandstone reservoir. *Energy Procedia*, in press.

Law, D.H.S. & Bachu, S. 1996. Hydrogeological and numerical analysis of CO₂ disposal in deep aquifers in the Alberta Sedimentary Basin, *Energy Conversion and Management*, **37** (6-8), 1167-1174.

Lindeberg, E.; Wessel-Berg, D. Vertical Convection in an Aquifer Column under a Gas Cap of CO₂. *Energy Conversion Management*, 1997, **38**, S229.

Mohitpour, M., H. Golshan, A. Murray. 2003. Pipeline Design & Construction - A practical approach. The American Society of Mechanical Engineers, Three Park Avenue, New York, United States.

Pedersen, S., Sjolvoll, M. and Fostas, B. Flue gas degradation of amines. Presented at the IEA GHG workshop in Oslo, February, 2010. Available at (14/12/2010): http://www.ieaghg.org/docs/General_Docs/Env_Impacts/3%20Flue%20Gas%20Degradation_Statoil_Pedersen.pdf

Pruess, K., Xu, T., Apps, J., García, J., 2003. Numerical modeling of aquifer disposal of CO₂. *SPE (Society of Petroleum Engineers) Journal*, p. 49-59, March (2003).

Seiersten, M., Material selection for separation, transportation and disposal of CO₂, *Proceedings Corrosion 2001*, National Association of Corrosion Engineers, paper 01042.

Soloman, S. & Flach, T., 2010. Carbon dioxide injection processes and technology, in Mercedes Maroto-Valer, M., *Developments and innovation in carbon dioxide capture and storage technology. Volume 1: Carbon dioxide capture, transport and industrial applications. Woodhead Publishing Series in Energy*, 2010.

Steefel, C.I. (2001). GIMRT, version 1.2: Software for modeling multicomponent, multidimensional reactive transport. User's Guide, UCRL-MA-143182. Livermore, California: Lawrence Livermore National Laboratory.

Steefel, C.I. and MacQuarrie, K.T.B. (1996). Approaches to modeling reactive transport in porous media. In *Reactive Transport in Porous Media* (P.C. Lichtner, C.I. Steefel, and E.H. Oelkers, eds.), *Rev. Mineral.* **34**, 83-125.

UCL, 2010. Dedicated CCS legislation. Available at (21/12/2010): <http://www.ucl.ac.uk/cclp/ccsdedlegstorage.php#storage>

Van der Meer, L.G.H., 1995. The CO₂ storage efficiency of aquifers. *Energy conversion and management*, **36** (6-9), 513-518.

Visser, de, E., Hendriks, C., Barrio, M., Molnvik, M., de Koeijer, G., Liljemark, S. and le Gallo, Y., 2008. Dynamis CO₂ quality recommendations. *International journal of greenhouse gas control* **2**, 478-484.

Visser, de, E., Hendriks, C., de Koeijer, G., Barrio, M., Liljemark, S., Brown, A. and Austegard, A. 2007. Dynamis CO₂ quality recommendations. Available at (21/10/2010):

http://www.dynamis-hypogen.com/publications/Deliverables/Year_2/D3-1-3%20DYNAMIS%20CO2%20quality%20recommendations.pdf

Wells, A. W.; Diehl, J. R.; Bromhal, G.; Strazisar, B. R.; Wilson, T. H.; White, C. M. 2007. The use of tracers to assess leakage from the sequestration of CO₂ in a depleted oil reservoir, New Mexico, USA. *Applied Geochemistry*, **22**, 996.

Xu, T., Apps, J.A., Pruess, K., 2004. Numerical simulation to study mineral trapping for CO₂ disposal in deep aquifers. *Appl. Geochem.* 19, 917-936.

Xu T, Apps JA Pruess K and Yamamoto H., 2007. Numerical modelling of injection and mineral trapping of CO₂ with H₂S and SO₂ in a sandstone formation. *Chemical Geology* 242, 319-346.

Yan, J., Anheden, M. and Bernstone, C. (2009). Impacts of non-condensable components on CO₂ compression/purification, pipeline transport and geological storage. *Proceedings of the 1st IEA Oxyfuel Combustion Conference*. Cottbus, September 8-11, 2009.